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Poojashri Ravindra Naik<sup>1</sup>, Vinod Alurdoddi Rajashekara<sup>2</sup>, Rajalakshmi Mudbidre<sup>3</sup>

# Estimation of Natural Uranium and Its Risk-Assessment in Groundwater of Bangalore Urban District of Karnataka, India

Abstract: In this study, 56 groundwater samples were taken from diverse sources in Bangalore Urban district during the pre-monsoon and post-monsoon seasons to measure the uranium concentration and its correlation with different waterguality parameters. The uranium concentration varied from 0.94–98.79 µg/L during the pre-monsoon season and from 1.38–96.52 µg/L during the postmonsoon season. Except for a few readings, all were within the safe limit of 60 µg/L as prescribed by India's Atomic Energy Regulatory Board (AERB), Department of Atomic Energy (DAE). Owing to its slightly higher concentration, a study on the radiological and chemical risks that are caused due to the ingestion of uranium was assessed. Based on the radiological aspect, cancer mortality and its risks were assessed, wherein all of the samples were well within the acceptable limit of 10<sup>-4</sup>; therefore, consuming these water samples was radiologically safe. However, when the risk that was caused by chemical toxicity was assessed, a few samples exceeded the hazard quotient (HQ) value of more than 1, thus illustrating that individuals were vulnerable to chemical risk. This paper features assessments of uranium and its risks to public health in groundwater samples if it exceeded the safe limit. Additionally, it recognizes the value of periodically assessing and treating the area's drinking water sources.

Keywords: groundwater, Bangalore Urban district, uranium, correlation, risk-assessment

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R V College of Engineering, Department of Civil Engineering, Bengaluru, India; Visvesvaraya Technological University, Belagavi, India, email: poojanaik22@gmail.com (corresponding author),
https://orcid.org/0009-0000-6274-8004

<sup>&</sup>lt;sup>2</sup> R V College of Engineering, Department of Civil Engineering, Bengaluru, India, email: vinodar@rvce.edu.in, https://orcid.org/0000-0003-4333-8813

<sup>&</sup>lt;sup>3</sup> R V College of Engineering, Department of Chemical Engineering, Bengaluru, India, email: rajalakshmim@rvce.edu.in, https://orcid.org/0000-0002-9503-4216

# 1. Introduction

Earth's crust contains uranium (U) - a primordial radionuclide. In nature, U exists as a rare radioactive heavy metal. Its natural distribution varies from rock-like structures to mineral deposits of various forms [1]. It is found as an oxide  $(UO_2^{2+} - ura$ nyl) in its most stable oxidation state (VI) in groundwater. It is the most long-living radionuclide, which produces ions with +4 (UO<sub>2</sub> and  $U^{4+}$ ) and +6 (UO<sub>3</sub> and  $UO_2^{2+}$ ) oxidation states, of which U (+6) (the uranyl  $(UO_2^{2+})$  ion – uranium oxide) reportedly has a higher solubility and the ability to form stable complexes with a variety of organic and inorganic ligands [2]. Since uranium is a naturally occurring lithophilic element in groundwater, its level in groundwater is dependent on the local lithology, geomorphology, and other geological factors [3]. The occurrence and distribution of uranyl species in surface and subsurface water are controlled by the redox condition, pH, and CO<sub>2</sub> partial pressure [4, 5]. As U (+6) as a uranyl ( $UO_2^{2+}$ ) compound is more soluble, it enters and contaminates groundwater by leaching from rock structures. Meanwhile, the use of uranium-containing pesticides and insecticides on land result in surface water contamination [6, 7]. Due to leaching, the components of uranium that are present in pesticides/insecticides may reach groundwater by contaminating it with uranium [8]. Adding to this, various anthropogenic sources such as mining could also contribute to high uranium concentrations in groundwater [9, 10].

There have also been reports of uranium contamination in drinking water from a few researchers worldwide. According to studies by Drury et al. [11], the ranges of the uranium content in the studied U.S. drinking water were 0.01–652, 0.11–640, and 0.02–6.99  $\mu$ g/L, respectively [11, 12]. Similarly, surface water in Ohio (USA), Datong Basin (China), and Singai (Egypt – uranium mineralized) witnessed uranium concentrations of 0.3–3.9, 1.2–16, and 600–1130  $\mu$ g/L, respectively [13]. Ribrira da Parithana of Portugal recorded a uranium concentration of 7.7–48.6  $\mu$ g/L [14], and the rivers of China recorded concentrations that ranged 3.85–7.57  $\mu$ g/L [15].

Groundwater has been discovered to contain higher levels of naturally occurring uranium in smaller geographic areas around India. Major uranium discoveries in India have occurred at Jaduguda (Jharkhand); there have also been occurrences of uranium found in the Cuddapah basin of Andhra Pradesh [16]. These include Lambapur-Peddagattu, Chitrial, Kuppunuru, Tumallapalle, and Rachakuntapalle, which have contributed significantly toward the uranium reserve base of India [17]. Other states like Rajasthan, Karnataka, and Chhattisgarh also have some major deposits of uranium.

The AERB (Atomic Energy Regulatory Board, Department of Atomic Energy, 2004) strictly regards a uranium concentration of 60  $\mu$ g/L as a safe limit in drinking water. To determine the uranium content in various locations in India, numerous researchers have undertaken investigations. In line with this, Sandeep et al. reported uranium concentrations of 6.37 and 43.31  $\mu$ g/L in the Bhiwani and Sirsa districts of Western Haryana, respectively [18]. Eastern Harayana reported uranium

concentrations that varied from 9.1 to 155.1  $\mu$ g/L [19]. Sharma and Singh [20] revealed uranium concentrations that varied from 0.13 to 1340  $\mu$ g/L in the Mansa district of the Malwa region. Almost 81% of their samples exceeded the guideline value of 60  $\mu$ g/L (as prescribed by the AERB). Santosh et al. carried out studies in the Durg district of Jharkhand, which witnessed a maximum uranium concentration of 45.7  $\mu$ g/L [21]. Kedar (Mandakini) Valley of Uttarkhand, Bagjata of Jharkhand, and the Rajnandgaon district of Chattisgarh witnessed varying uranium concentrations of 0.02–63.7, 0.5–11.2, and 0.5–99.0  $\mu$ g/L, respectively. Meher et al. [22] reported an average uranium concentration of 2.75  $\mu$ g/L in the Alakananda River and 1.86  $\mu$ g/L at the River Ganga. Kashyap and Ghosh [23] mentioned concentrations of 0.031 to 140.1  $\mu$ g/L in the Korba district, Chhatisgarh. The Bagjata and Banduhurang districts of Jharkhand witnessed concentrations of 0.5–11.2 and 0.5–27.5  $\mu$ g/L, respectively [8, 21].

The southern region of India has been the subject of numerous inquiries. Central Tamil Nadu reported uranium concentrations of 11.78-68.28 µg/L, which was due to the dominance of lithological granitic structures like granite, quartzite, fissile hornblende, biotite, and gneiss [22, 23]. The Palakkad, Thrissur, Kottayam, and Ernakulam districts of Kerala witnessed 0.5-12.54 µg/L uranium concentrations [24], whereas salem district of tamil nadu and central part of tamil nadu was reported with 0.01–385.40 µg/L and 0.4–68.28 µg/L respectively [25, 26]. The Nalgonda district, Vishakaptanam, and Tummalapalle of Andhra Pradesh have been reported to have varying uranium concentrations of uranium (0.2-68.0, 0.6-12.3, and 0.38-79.70 µg/L, respectively). From the above results that were reported in different parts of India, it can be presumed that groundwater is used as the sources of drinking water by 40% of the population in most locations. People residing in urban areas consume water with prior treatment; hence, the size of the exposed population is small. On the other hand, people residing in the rural areas depend on groundwater more often in the form of community or individual wells for drinking and agricultural purposes. Stating varying concentrations of uranium in different parts of India, it can be observed that the residents rely on groundwater from basement aquifers for drinking water supplies as well as domestic and agricultural use. So, it is a significant concern to assess elevated levels of uranium concentrations that are being exposed to the public [27].

When the public is exposed to elevated levels of uranium concentrations, it leads to health risks because of its chemical and radiological toxicity. The route of the exposure, the solubility of the particles, the contact time, and the method of the disposal are the main grounds for uranium morbidity. The uranium content and its mobility are controlled by pH, ORP (oxidation and reduction potential) and, complexing agents such as sulfates, fluorides, phosphates, carbonates, etc. in fluid systems [28, 29]. The presence of U (VI) in water (either ground or surface) has garnered significant attention; when someone is exposed to high levels of uranium, it could damage their kidneys and cause nephritis. Moreover, intakes of U above 60 ppb is known to cause acute kidney failure and is carcinogenic to human health [30, 31].

Considering the significance of determining uranium concentrations in the context of their impact on human health and the environment, an effort has been made in the current work to estimate the concentrations of uranium (U) in ground-water sources during the pre-monsoon (PRM) and post-monsoon (POM) season of the Bangalore Urban district, study its correlation with water-quality parameters, and estimate the radiological and chemical risks that can be attributed to uranium ingestion.

## 2. Geology of Study Area

The Bangalore Urban district is the most populous district in the Indian state of Karnataka, which is spread across 2174 km<sup>2</sup> and is located between the north latitudes of 12°39′32′′ and 13°14′13′′ and the east longitudes of 77°19′44′′ and 77°50′13′′; it is located on the Deccan plateau at a height of more than 900 m (3000 ft) above sea level. The Bangalore Urban district is surrounded by the Bangalore Rural district to the east and north, the Ramanagara District to the west, and the Krishnagiri District of Tamil Nadu to the south. The Bangalore Urban district has four taluks; i.e., Bangalore North, Bangalore South, Bangalore East, and Anekal Taluk (Fig. 1). The dominant rock structures that are spread across the Bangalore Urban District are granites, gneisses, and the peninsular gneissic complex (PGC) that is associated with Closepet and other granitic rocks (Fig. 2). The dominant soil types are red laterite, argillaceous, and from fine loamy to clay-like.



Fig. 1. Map of Karnataka and locations of sampling sites



Fig. 2. Geology of Karnataka map Data source: Department of Mines and Geology, Government of India

The Arkavathi River Basin, which makes up 46% of the district, is to the west. The Dakshina Pinakini River Basin, which makes up 54% of the district, is to the east. Numerous elements, including the fracture pattern, level of weathering, geomorphological configuration, and rainfall amount, influence groundwater circulation and the recharging of the aquifers. The resistivity tests revealed the existence of weathered (permeable) rock that extend down to a depth of 30 m. The major aquifer is located at a depth of 25–30 m. Even below 60 m, there are still aquifers [32, 33].

The study region is situated on the tonalitic biotite gneisses-dominated Archean gneiss complex of Peninsular India. Along with a few tanks and lift-irrigation schemes, groundwater is the district's main source of irrigation water [34]. The two main crops that are farmed in the district are paddy and ragi, and the supporting crops including maize, cereals, and groundnuts. The Precambrian migmatite, granodiorite, tonalite, and gneiss that underlie the Bangalore study region have local granitic intrusions. As reported by Sekhar et al. [35], the depth of the groundwater table is merely a few meters at a few sites in Central Bangalore; in other places, however, it can be as deep as 70 m. The geology and fracture network, seasonal monsoon recharge, current and historical abstraction, and (in the case of Bangalore) urban leakage all impact the groundwater levels. The key issues in the Bangalore Urban district are sewage contamination and industrial pollution; because of these, a high nitrate concentration can be witnessed in the subsurface areas. The over-exploitation of groundwater resources (due to the increased urbanization), and the excavations for metro train construction are the impacts that have caused the mentioned factors [36].

# 3. Materials and Methods

#### 3.1. Collection of Water Samples

A 6 km × 6 km grid map of the Bangalore Urban district was prepared using ArcView GIS software (Version 3). A total of 56 water samples (Fig. 1) were taken from the surrounding borewells and other drinking water sources from nearby sampling locations. Where a groundwater source was not available, surface water was collected as a sample source. Before collecting the water samples, they were kept running for 2–5 minutes before collecting them in pre-cleaned polyethylene bottles. Ex-situ parameters like pH, TDS, EC, temp and dissolved oxygen were analyzed at the sampling sites, while other water-quality parameters like nitrates, total hardness, phosphates, sulfates, chloride, and uranium were analyzed in the laboratory. Using the Global Positioning System (GPS), the locations of each sample were noted.

#### **3.2.** Uranium Estimations in Water Samples

The uranium concentrations that were present in the water samples were examined using an LF-2a LED Fluorimeter (manufactured by Quantalase Enterprises Pvt. Ltd., Indore, India). The instrument consisted of light-emitting diodes (LED) along with photo multiplier tubes (PMTs). When a water sample was placed in a quartz cuvette, a pulsed UV light was emitted at a 400 nm wavelength (which was passed through a suitable filter). The green fluorescence that was released due to the excitation of the uranium species at the applicable wavelength was measured by the photo multiplier tubes (PMTs) in this method. The uranium concentrations in the fluid samples specified the estimations of the fluorescence. Stable uranyl phosphate, a fluorescence-enhancing reagent, was produced by adding a sodium pyrophosphate solution [21]. Ortho-phosphoric acid (OPA) helped in maintaining the pH level of the reagent at 7 [16]. To verify the accuracy of the results, the laser fluorimeter (LF) was calibrated using a standard uranium solution of a known concentration, and reagent blanks were run with samples of distilled water [37]. A uranium stock solution of 100 µg/mL was used to prepare the working standard samples in order to calibrate and test the instrument's performance. To reduce errors during the sample analysis, an analytical balance and automated pipette were utilized. A sample of water was blended with 0.5 mL of 5% sodium pyrophosphate in a dry spotless cuvette (maintained at pH = 7), and the results were measured. Since the water samples were collected from several places with different chemical compositions, the matrix effect had to be avoided throughout the analysis by employing the standard addition procedure [38].

#### 3.3. Determination of Other Physico-Chemical Parameters

During the pre-monsoon and post-monsoon seasons, numerous physicochemical parameters were scientifically investigated, along with the assessments of uranium in the water samples from the Bangalore Urban district according to Table 1. The analysis of the water was carried out in line with standard methods. pH and TDS (total dissolved solids) were measured using pH and TDS probes. Nitrates (NO<sub>3</sub><sup>-</sup>), phosphates (PO<sub>4</sub><sup>3-</sup>) and sulfates (SO<sub>4</sub><sup>2-</sup>) were measured using a portable nitrate photometer (HI 96786), a phosphate HR-checker (HI 717), and a portable Sulfate photometer (HI 94751) supplied by Hanna Instruments India Pvt. Ltd. The total hardness (TH) in the water was calculated using the EDTA titrimetric technique. Using an ion-selective electrode (ISE) (HI 5522) that was supplied by Hanna Instruments India Pvt. Ltd., fluoride (F<sup>-</sup>) was detected. Chloride (Cl<sup>-</sup>) was determined using the argentometric method, and uranium (U) was analyzed using the LF 2a LED fluorimeter model supplied by Quantalase Enterprises Pvt. Ltd.

Sample	Paramotor		Para	meter va	lues	Mathad fallowed
no.	1 af affilieter	WIIO (2011)/ DIS/ AEKD	value	POM	PRM	Method followed
			min	6.9	7.1	
1	рН	6.5-8.5	max	8.0	8.3	pH probe
			average	7.38	7.59	
	total dissolved		min	125	132	
2	solids (TDS)	500 (BIS)	max	1150	1040	TDS probe
	[mg/L]		average	337.39	609.36	
			min	1.5	1.3	
3	nitrate	50	max	81	22	nitrate photometer
	[IIIg/L]		average	13.39	12.93	
			min	15	65	
4	total hardness	600 (BIS)	max	755	520	EDTA titrimetric
	[IIIg/L]		average	360.00	370.88	memou
			min	0.2	0.3	
5	phosphate	none	max	2.9	2.5	phosphate checker
	[IIIg/L]		average	1.20	1.36	
			min	2	12	
6	sulfate	500	max	156	120	sulfate photometer
	[mg/L]		average	59.45	90.71	-
			min	0.25	0.72	
7	fluoride	1.5 (BIS)	max	1.98	1.90	ion selective
	[mg/L]		average	1.05	1.40	electrode
			min	5	0	
8	chloride	250 (BIS)	max	285.91	84.00	argentometric
	[mg/L]		average	55.75	44.00	method
			min	1.38	0.94	
9	uranium	60 (AERB)	max	96.52	98.79	LED fluorimeter
	[µg/ւ]		average	24.57	23.52	LF Za

Table 1. Analysis of various physico-chemical parameters of water

#### 3.4. Correlation Analysis of Uranium with Water-Quality Parameters

After analyzing the uranium along with the other water-quality parameters, its significance with the different water-quality parameters was studied further [34]. The association between uranium and other water-quality parameters was assessed using the Number Cruncher Statistical System (NCSS). To determine the correlation coefficients, the coefficients of many sets of parameters were computed and used to build a correlation matrix [39]. The findings of the uranium tests were connected with the various indicators of water quality.

#### 3.5. Human Risk-Assessment Due to Uranium Ingestion

The uranium concentrations were higher than the permitted limit at a few sample sites. Assessing the health impact on the general public becomes vital when people are exposed to higher uranium concentrations in their drinking water sources. When uranium is exposed to the environment, it affects human health; this can be classified as a radiation risk that is caused by the radiation of the isotopes of any uranium that is present in water and the chemical risk that is caused by its toxicity (as it is a radioactive element). The radiological and chemical risks that were assessed for the collected water samples will be discussed further [40].

## 4. Results and Discussion

## 4.1. Estimation of Uranium with Different Water-Quality Parameters

In the current investigation, 56 water samples were collected during the pre-monsoon and post-monsoon seasons. The uranium concentrations in the pre-monsoon season varied from 0.94 to 98.79  $\mu$ g/L, and in the post-monsoon season, these varied from 1.38 to 96.52  $\mu$ g/L. A few samples from both the pre-monsoon and post-monsoon seasons exceeded the AERB limit; Kadubeesanahalli (Sample No. 11) and Kaggalipura (Sample No. 32) witnessed higher uranium concentrations (76.85 and 98.79  $\mu$ g/L, respectively) during the pre-monsoon season, whereas Dibburu (Sample No. 14) and Channenahalli (Sample No. 42) witnessed 96.57 and 84.99  $\mu$ g/L, respectively, during the post-monsoon season (Fig. 3, Table 2).

The uranium concentrations that were discovered in the groundwater in this study were comparable to those that were discovered in previous recent investigations [6, 36]. The minimum uranium concentrations in the Bangalore Urban district were 0.94 (PRM) and 1.38  $\mu$ g/L (POM), whereas the maximum concentrations were 98.79 (PRM) and 96.52  $\mu$ g/L (POM) (Table 1). It could also be observed from the study that the uranium concentrations varied within shorter distances from the populations; this was in line with the similar tendencies that were mentioned by Nagaiah et al. [41]. This made it so that the distributions of the naturally occurring uranium in the groundwaters reflected the local environments [42].

As reported by Lapworth et al. [34] urban leakage in Bangalore led to a low-pH and low-bicarbonate groundwater hydrochemistry, thus reducing uranium's mobility and altering uranium's speciation. The local geology and physico-chemical environment may be responsible for the differences in the uranium concentrations at various distances. Leaching is more effective when bicarbonate is formed while water is seeping through the soil. This may be one of the mechanisms that underlay the high levels of uranium in the local groundwaters at a few sampling sites, but other explanations remain conceivable. Kadubeesanahalli is an IT hub, and Brigade Meadows (in Kaggalipura) is a group of high-rise apartments. The reason behind the slighter higher concentrations could have resulted from the overlying of the soils (one above the other). The leakage of septic tanks, the infiltration of wastewater, and (primarily) the decline of the groundwater table might have triggered geogenic contaminants with uranium mobilization to the groundwater from its sediments. The flow of the groundwater along with any geogenic contaminants through aquifer zones might have reached the bore wells, which might have been a contributing factor to the rises in the uranium concentrations [43].



Fig. 3. Variations of uranium contents during pre-monsoon and post-monsoon seasons in Bangalore Urban district

The majority of the groundwater samples in the study areas were of the alkaline type; pH levels varied from 6.9 to 8 during the post-monsoon season and from 7.1 to 8.3 during the pre-monsoon season (Fig. 4) [44]. Meanwhile, the total dissolved solids varied from 125 to 1150 mg/L during the post-monsoon season and from 132 to 1040 mg/L during the pre-monsoon season (Fig. 5). The samples that were collected during the post-monsoon season noted a high TDS value of 1150 mg/L; during the pre-monsoon season, they were up to 1040 mg/L. Because too many minerals

were dissolved, the TDS of certain water samples exceeded the prescribed limit. One explanation for these increases in the TDS contents could have been due to increases in urban runoff and the use of agricultural fertilizers [45, 46]. The phosphate concentrations ranged from 0.2 to 2.9 mg/L in the post-monsoon and from 0.3 to 2.5 mg/L in the pre-monsoon seasons (Fig. 6) [41–44]. The sulfate concentrations varied from 12 to 120 mg/L during the pre-monsoon and from 2 to 120 mg/L during the post-monsoon seasons (Fig. 7). The sulfate concentrations were well within 200 mg/L (as per the BIS standards). The nitrate concentrations varied from 1.3 to 22 mg/L during the pre-monsoon and from 1.5 to 81 mg/L during the post-monsoon seasons (Fig. 8). Overall, our data sets showed no discernible correlation between the uranium and nitrates, which ruled out the possibility that fertilizers were the primary sources of the uranium. The nitrate levels were well within BIS recommendations.



Fig. 4. Variations of uranium with pH during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district



Fig. 5. Variations of TDS along with uranium during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district



Fig. 6. Variations of phosphate with uranium during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district



Fig. 7. Uranium variations with sulfate during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district



Fig. 8. Variations of uranium with nitrates during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district

The calcium hardness (CaH) varied from 30 to 460 mg/L during the pre-monsoon and from 5 to 495 mg/L during the post-monsoon seasons (Fig. 9). The magnesium hardness (MgH) varied from 0 to 320 mg/L during the pre-monsoon and from 0 to 625 mg/L during the post-monsoon seasons (Fig. 10).



Fig. 9. Variations of calcium hardness with uranium during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district



Fig. 10. Variations of magnesium hardness with uranium during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district

The fluoride concentrations varied from 0.25 to 1.98 mg/L during the postmonsoon and from 0.72 to 1.9 mg/L during the pre-monsoon seasons (Fig. 11). They were within the prescribed limit; these might exceed the limit if no proper monitoring is followed; the risk of dental fluorosis increases with fluoride concentration, while the risk of skeletal fluorosis increases at higher concentrations. The high fluoride contents in the groundwater may have been caused by increases in infiltration via weathered rocks, the use of phosphatic fertilizers, and other industrial activities [47]. Hence, the sources of groundwater supplies to villages with high concentrations of fluoride must be treated with defluoridation techniques.



Fig. 11. Variations of uranium with fluoride during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district

As shown in Figure 12, the chloride concentrations varied from 0 to 84 mg/L during the pre-monsoon and from 5.00 to 285.91 mg/L during the post-monsoon seasons.



Fig. 12. Variations of uranium with chloride during pre-monsoon (a) and post-monsoon (b) seasons in Bangalore Urban district

Figure 13 illustrates the schematic ion ratios of  $Ca^{2+}/Mg^{2+}$  that are used to trace the sources of calcium and magnesium in groundwater. According to recently conducted research that was carried out by Srinivasamoorthy et al. [48], it indicated that, if the ratio of  $Ca^{2+}/Mg^{2+} > 2$ , this infers that silicate minerals have an impact by adding calcium and magnesium ions to the groundwater. Along similar lines, Ajay Kumar et al. [4] inferred that, if the molar ratio of  $Ca^{2+}/Mg^{2+} > 1$ , there might have been either a high decomposition of calcium-bearing minerals like calcite, anhydrite, and gypsum or a dissolution of calcium-containing silicate minerals (igneous – including granite rocks) during the geochemical processes. As compared to our study, around 50% of the samples had levels that were greater than 2 in both the pre-monsoon and post-monsoon seasons. The Closepet granites and gneiss rock structures contained silica, which might have been the reason for the increase in the ratio of  $Ca^{2+}/Mg^{2+} > 2$ . In the post-monsoon season, fewer than 25% of the samples fell below the 1-ratio line, which means that the  $Ca^{2+}$  might have precipitated as  $CaCO_3$  due to the ion-exchange process. As the magnesium composition is comparatively less in the parent rock of the Bangalore Urban district, it has been less-witnessed.



Fig. 13. Relationship of major ions: scatter diagram of Ca<sup>2+</sup>/Mg<sup>2+</sup> during pre-monsoon and post-monsoon seasons

However, recent studies that were conducted by Lapworth et al. [34] indicated that, urban leakage in Bangalore caused low pH and low bicarbonate groundwater hydrochemistry, which changed uranium speciation and decreased uranium mobility. Nevertheless, identical uranium concentrations in comparable environments have been recorded in India – both in the crystalline basement and the alluvial soil aquifer systems. As a result, the study highlighted the uranium concentrations in various villages inside the Bangalore Urban zone [48].

Table 2 illustrates details of the sampling locations and the values of the different water-quality parameters: pH, total dissolved solids (TDS), total hardness, nitrates, phosphates, sulfates, chloride, fluoride, and uranium concentrations. Table 2. Details of sampling location along with values of different water quality parameters

on details	pł	H	TL [mg	JL]	Nitra [mg	ates /L]	Tol hard [mg	tal ness ;/L]	Phosp [mg	hates /L]	Sulf: [mg	ates /L]	Chlo [mg	ride /L]	Fluo [mg	ride 5/L]	Urani [µg/	tum TL]
	PRM	POM	PRM	POM	PRM	POM	PRM	POM	PRM	POM	PRM	POM	PRM	POM	PRM	POM	PRM	POM
illi	7.6	7.3	926	532	12	5.2	405	485	1	1.3	108	82	20	50	0.83	0.56	7.23	16.5
	7.7	7.15	606	265	15	6.8	480	320	1.2	1.5	112	86	55	30	1.39	2.71	5.62	14.3
ar	7.4	7.28	590	215	12	9.2	300	115	0.8	1.9	42	87	50	15	3.82	06.0	7.42	18.3
ar	7.3	7.3	902	242	13	1.5	350	125	1	1.2	35	82	30	80	3.07	0.25	8.01	1.76
har	7.4	7.62	461	425	12	12.5	400	125	0.9	2.5	78	78	45	50	2.12	0.17	21.3	15.5
t	7.3	7.5	450	525	15	14.5	500	130	2.1	2.25	59	80	59.	50	2.3	0.25	16.0	17.9
	7.6	7.81	625	321	10	13.2	255	120	2.2	1.5	60	68	55	60	1.85	0.13	18.2	13.7
ura	7.5	7.62	766	285	6.9	11	280	130	2	1.4	56	72	70	60	2	0.98	34.8	4.44
	7.2	7.25	132	402	1.5	5.6	105	15	1.2	1.2	15	81	ß	35	8.89	0.39	1.52	1.85
tus s	7.2	7.62	584	560	1.3	12.2	410	280	2.5	1.6	12	83	5	25	7.3	0.28	14.9	42.1
nahalli	7.5	7.8	403	580	5.8	5.2	65	225	2.1	0.8	59	6	I	125	4.52	0.81	76.8	13.4
	7.7	7.1	144	1070	8.8	2.9	190	475	0.8	1.3	30	89	25	50	3.54	0.91	29.1	9.3
i	7.7	7.4	700	182	13	12	405	600	1.8	1.2	118	62	55	60	2.84	0.24	4.98	19.5
	7.7	7.5	740	160	14	12	420	250	1.72	1.2	118	27	50	35	2.36	3.2	9.06	96.5
	7.7	7.4	798	810	14	27.6	390	525	1.7	0.7	112	67	35	55	2.36	0.24	10.2	6.07
E E	7.7	7.3	142	165	18	4.8	425	435	1.23	0.9	112	51	55	65	1.79	4.25	11.2	8.64
illi	7.8	7.6	802	205	21	12	495	385	1.7	1.3	118	43	60	35	1.02	0.28	10.9	13.7
ahalli	7.8	7.6	560	211	18	2.4	310	405	1.2	1.3	112	34	45	50	1.82	2.68	4.75	40.2

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ium /L]	POM	45.9	51.3	48.6	67.4	26.3	15.97	7.35	3.97	23.2	66.7	27.5	4.57	1.38	35.7	29.7	18.6	3.92	
Uran [µg	PRM	8.95	18.5	11.2	28.2	0.94	8.77	3.45	11.3	5.37	61.3	58.9	9.84	5.84	98.7	16.0	25.8	34.6	
ride /L]	POM	1.76	1.66	0.57	0.52	0.21	0.56	1.58	2.44	2.22	0.98	0.83	1.52	0.92	0.89	1.51	0.54	1.62	
Fluo: [mg	PRM	1.52	1.78	3.34	4.45	5.4	4.7	4.2	3.69	1.42	0.72	0.82	0.29	0.80	1.4	1.2	1.62	1.3	
ride /L]	POM	55	65	70	70	Ŋ	65	30	55	25	60	10	30	80	285	55	65	60	
Chlo [mg	PRM	59	54	49	45	85	45	40	40	15	45	25	55	70	15	ы	40	55	
ates /L]	POM	42	35	32	32	œ	81	85	82	81	82	85	86	151	120	126	156	61	
Sulf: [mg	PRM	70	100	80	112	112	112	104	92	102	90	1	112	42	72	98	100	120	
hates /L]	POM	2.3	2.4	2.8	1.8	0.4	0.5	2.9	2.8	2.5	1.2	1.6	1.5	1.2	0.8	0.9	1.5	0.3	
Phosp [mg	PRM	1.5	1.3	1.6	1.2	1.8	0.9	0.9	0.8	0.9	0.6		1.2	1	1.8	1.8	1.5	0.8	
al ness /L]	POM	560	510	450	440	615	360	280	55	340	450	610	320	115	390	445	555	370	
Tot hardı [mg	PRM	412	520	510	430	375	245	300	395	340	370	490	480	190	415	465	432	455	
ates /L]	POM	5.5	12.5	11.2	12	12	6.1	12	17	17	12.3	5.6	6.8	6.2	27.3	28	56	12	
Nitra [mg	PRM	15	14	10	21	22	8	12	14	14	9.5	12	15	14	21	12	13	14	
S /L]	POM	325	423	250	485	589	336	562	625	325	250	450	265	245	1150	1120	562	140	
[mg	PRM	680	720	830	989	1040	333	568	568	317	712	683	606	453	644	152	250	820	
Ŧ	POM	7.5	7.6	7.7	7.2	7.2	7.6	7.6	7.52	7.4	7.7	7.25	7.15	7.3	7.2	7.2	7.2	7.1	
pŀ	PRM	7.8	7.3	7.5	7.4	7.2	7.9	7.1	7.3	7.6	7.9	7.8	7.7	7.2	7.3	7.4	7.5	7.2	
Location details		Vidyaranyapura	Jalahalli	Abbigere	Byadarahalli	Kuntanahalli	Chikkagubbi	Aduru	Maralakunte	Budigere	Ullal	Padmana- bhanagar	R R Nagar	BHEL Layout	Kaggalipura	Talaghattapura	Anekal	Gottigere	Banorchatta
Sample no.		19	20	21	22	23	24	25	26	27	28	29	30	31	32	33	34	35	

Table 2. cont.

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10.6	14.4	16.3	3.92	8.68	84.9	9.71	6.02	33.0	12.4	50.6	70.1	20.2	6.28	24.4	29.3	43.8	33.2	23.1	33.0
41.2	56.1	46.8	34.6	23.5	2.2	25.5	3.25	58.3	36.1	17.8	33.4127	38.8	16.8	2.17	14.3	15.0	57.5	6.71	58.3
0.96	0.81	0.81	0.62	0.65	0.86	0.58	0.61	0.59	0.84	0.68	0.52	0.96	1.06	1.02	1.04	1.53	1.32	1.39	1.85
1.8	1.4	1.2	1.3	1.3	1.2	1.3	1.32	1.3	1.49	1.2	1.25	1.6	1.7	1.2	1.48	1.2	1.2	1.82	1.2
58	55	60	60	50	65	50	55	40	55	60	65	50	50	50	65	55	55	55	ы
55	49	45	55	45	55	65	25	60	50	45	55	65	55	35	40	25	65	60	ы
22	2	17	61	23	4	32	65	32	7	17	4	32	34	150	44	48	28	56	110
90	100	113	120	111	113	110	100	112	100	113	113	110	100	110	100	100	87	101	112
0.8	0.2	0.8	0.3	0.3	0.6	1.4	0.9	1.1	0.2	0.8	0.6	1.4	0.4	0.2	0.6	1.6	1.4	0.7	0.7
1.9	1.4	1.2	0.8	1.2	1.8	1.8	1.4	1.72	1.4	1.2	1.8	1.8	1.6	0.3	1.2	1.2	0.8	0.8	1.7
425	595	285	370	225	60	755	255	440	595	285	60	755	655	295	590	275	225	480	310
350	370	350	455	285	385	425	380	375	370	350	385	425.00	305	390	355	375	445	435	265
15	7.5	7.2	12	16.6	12	20.1	22.4	4.6	7.5	7.2	12	20.1	81	12	15.1	æ	7.6	10	5.5
15	~	14	14	9	22	12	18	18	~	14	22	12	4	×	15	4	14	18	22
135	152	182	140	125	358	487	458	205	152	182	358	487	156	185	145	162	178	252	424
260	417	814	820	565	645	863	389	597	417	814	645	863	568	631	550	488	917	522	481
7.2	7.62	7.32	7.1	7.25	7.4	7.3	7.6	7.5	7.62	7.32	7.4	7.3	7.3	7.5	7.3	6.9	7.6	~	7.1
7.3	7.6	8.2	7.2	8.2	7.3	7.5	7.6	7.5	7.6	8.2	7.3	7.5	æ	8.3	8.1	7.9	7.2	7.7	7.5
Singasandra	Harapanahalli	Chandapura	Chandrashek- harapura	Golahalli	Chennenahalli	Dodda Alada Mara	Kitthanahalli	Gejjigadahalli	Attibele	Sarjapura	Chikka Thirupati Rd	Bukkasagara	Vaddarapalya	Vinayaka Nagar, Anekal	Yelahanka	Husmaranahalli	Kanakapura	Gerahalli	Haniyur
37	38	39	40	41	42	43	44	45	46	47	48	49	50	51	52	53	54	55	56

#### 4.2. Analysis of Uranium with Different Water-Quality Parameters

Utilizing the Statistical Package of Number Cruncher Statistical System (NCSS) program, it was possible to determine whether the water-quality parameters (n = 9) and uranium concentrations were correlated. A Pearson's correlation matrix approach was used, as the data was normally distributed.

Table 3 lists the correlation matrix that the software extracted during the pre-monsoon season. During this season, uranium showed weakly correlated positive levels with TDS (0.06), nitrates (0.05), and phosphates (0.16). Due to the extensive abstraction of the groundwater for household and industrial purposes, it exhibited considerable seasonal variability [35]. As reported, the Bangalore samples' hydrochemistry demonstrated low pH levels, low bicarbonate, decreased U mobility, lower occurrence, unique U speciation, and mineral saturation. Therefore, no such direct correlation could be observed in either the pre-monsoon or post-monsoon seasons (Table 4). The total hardness witnessed a positive correlation with TDS, fluoride, chloride and sulfates; this showed that the primary source of the cations was the weathering of the bedrock. Phosphates witnessed weakly correlated positive levels with fluoride, chloride, and nitrates. As mentioned in [34] samples that were drawn from Bangalore city also witnessed low phosphorous contents. Sulfates were found to have weak positive correlations with the pH levels, total dissolved solids, fluoride, chloride, and nitrates. Nitrates witnessed weakly positive correlations with TDS, fluoride, and chloride. Chloride had a positive correlation with total dissolved solids, while fluoride had a weak correlation with total dissolved solids. Apart from the above-mentioned positive correlations, the rest were negatively correlated. Referring to the table that was computed by Keith S. Taber, the alpha values were characterized as follows: excellent (0.93–0.94), strong (0.91–0.93), dependable (0.84–0.90), robust (0.81), reasonably high (0.76–0.95), high (0.73–0.95), good (0.71–0.91), relatively high (0.70–0.77), slightly low (0.68), reasonable (0.67–0.87), adequate (0.64–0.85), moderate (0.61–0.65), satisfactory (0.58–0.97), acceptable (0.45–0.98), sufficient (0.45–0.96), not satisfactory (0.4–0.55), and low (0.11). The Cronbach's alpha score of Pearson's matrix that was obtained during the pre-monsoon season was 0.3781, which was said to be low [49, 50].

Table 4 lists the correlation information during the post-monsoon season that the software extracted. During the pre-monsoon season, uranium showed a weak positive correlation with the pH levels (0.066), chloride (0.038), nitrates (0.062), and phosphates (0.08). The total hardness had weak positive correlations with the pH levels, chloride, nitrates, and phosphates. Phosphates witnessed weak positive correlations with the pH levels, TDS, fluoride, and sulfates. As mentioned in [34], samples that were drawn from Bangalore city also witnessed low phosphorous contents. Sulfates found weak positive correlations with TDS, fluoride, chloride, and nitrates. Chloride had weak positive correlations with the pH levels, total dissolved solids, and fluoride. Fluoride showed weak positive correlations with the total dissolved solids and pH levels. As the water-quality parameters had weak positive correlations, it can be inferred from the results that each parameter functioned on its own and was independent of the others. Apart from the aforementioned weak positive correlations, the rest were negatively correlated. The Cronbach's alpha score of Pearson's matrix that was obtained during the pre-monsoon season was 0.1939, which was said to be less reliable.

	рН	TDS	F-	Cl-	NO <sub>3</sub>	SO <sub>4</sub> <sup>2-</sup>	PO <sub>4</sub> <sup>3-</sup>	TH	U
pH	1.00								
TDS	-0.05	1.00							
F-	-0.29	0.15	1.00						
Cl-	-0.06	0.41	-0.07	1.00					
NO <sub>3</sub> -	-0.29	0.25	0.41	0.28	1.00				
SO4 2-	0.22	0.34	0.09	0.42	0.26	1.00			
PO <sub>4</sub> <sup>3-</sup>	-0.26	-0.01	0.34	0.02	0.05	-0.10	1.00		
TH	-0.11	0.41	0.26	0.36	0.35	0.48	-0.01	1.00	
U	-0.12	0.07	-0.06	-0.13	0.00	-0.04	0.17	-0.07	1.00

Table 3. Uranium correlation matrix for various pre-monsoon water-quality parameters

Cronbach's alpha – 0.3781.

TDS,  $F^-$ ,  $Cl^-$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $PO_4^{3-}$ , and TH were measured in mg/L; U in  $\mu$ g/L.

Table 4. Uranium correlation matrix for various post-monsoon water-quality parameters

	рН	TDS	F-	Cl-	NO <sub>3</sub>	SO <sub>4</sub> <sup>2-</sup>	PO <sub>4</sub> <sup>3-</sup>	TH	U
pН	1.00								
TDS	-0.02	1.00							
F-	0.22	0.14	1.00						
Cl-	0.05	0.35	0.05	1.00					
NO <sub>3</sub> -	-0.11	0.11	0.09	0.10	1.00				
SO4 2-	-0.08	0.36	0.36	0.09	0.07	1.00			
PO <sub>4</sub> <sup>3-</sup>	0.21	0.24	0.17	-0.12	-0.13	0.15	1.00		
TH	0.07	-0.06	-0.03	0.04	0.06	-0.30	0.08	1.00	
U	0.07	-0.06	-0.03	0.04	0.06	-0.30	0.08	1.00	1.00

Cronbach's alpha – 0.1939.

TDS, F<sup>-</sup>, Cl<sup>-</sup>, NO<sup>-</sup><sub>3'</sub> SO<sup>2-</sup><sub>4</sub>, PO<sup>3-</sup><sub>4</sub>, and TH were measured in mg/L; U in  $\mu$ g/L.

#### 4.3. Risk-Assessment of Human Health

Since India's groundwater supplies the majority of its drinking and agricultural water, the occurrence of high groundwater U concentrations could pose a risk to the general public's health. In line with the values that were obtained on uranium concentrations, it was inferred that its concentration exceeded the prescribed limit in a few samples. The current study was extended to evaluate excess cancer risk based on the estimated uranium values. The radiological and chemical risks will now be addressed in detail [40].

#### **Radiological Risk-Assessment**

The purpose of a radiological risk-assessment is to identify the radiation threats that may arise from the presence of different types of ionizing radiation by the various chemical substances that are present in water. These hazards are rarely assessed in drinking water, as this has less of a public health significance. The excess cancer risk (also called radiological risk) that can be attributed to uranium concentration is mentioned in Equations (1) and (2):

$$RR = U \cdot RF \tag{1}$$

where: RR – radiological risk, U – uranium [Bq/L], RF – risk factor.

The factors in Equation (1) are as follows:

$$U = UC \cdot CF \tag{2}$$

where: UC – uranium concentration  $[\mu g/L]$ , CF – conversion factor (0.025 Bq/ $\mu g$ );

$$RF = RC \cdot WIR \cdot D \tag{3}$$

where: RC – risk coefficient (cancer mortality risk/cancer morbidity risk), WIR – water ingestion rate [L/day], D – duration of total exposure [days].

Here, RF is the risk factor that is mentioned in Equation (3), the water ingestion rate is considered to be 4.05 L/day, and the average Indian life expectancy is considered to be 70.2 years (Press Information Bureau Government of India). Mortality represents the incidence of fatal cancer, and morbidity highlights the incidences of all cancers (fatal and non-fatal). Despite being a radioactive element without a stable isotope, U-238 is the most common naturally occurring form of uranium (>99.2%), and it is also the least radioactively dangerous. The cancer mortality and morbidity risks are featured in Tables 5 and 6 during the pre-monsoon and post-monsoon seasons. The cancer mortality and morbidity risks were calculated by considering the cancer mortality coefficient to be  $1.19\cdot10^{-9}$ /Bq and the cancer morbidity coefficient to be  $1.84\cdot10^{-9}$ /Bq [20, 40, 51]. For the Bangalore Urban district, the cancer mortality risk ranged from  $3.96\cdot10^{-8}$  to  $4.06\cdot10^{-6}$  during the premonsoon season (with an average of  $9.918\cdot10^{-7}$ ) and  $5.81\cdot10^{-8}$  to  $4.07\cdot10^{-6}$  during the post-monsoon period (with an average of  $1.07\cdot10^{-6}$ ). Bangalore's cancer morbidity risk varied from  $6.129 \cdot 10^{-8}$  to  $6.44 \cdot 10^{-6}$  during the pre-monsoon season (with an average of  $1.53 \cdot 10^{-6}$ ) and  $8.9 \cdot 10^{-6}$  to  $6.2 \cdot 10^{-6}$  during the post-monsoon period (with an average of  $1.6 \cdot 10^{-6}$ ). From the obtained results, it can be observed that all of the values of the cancer mortality and morbidity risks were well within the acceptable limit of  $10^{-4}$ . Consuming the water samples is, therefore, radiologically safe [20].

Table 5. Concentration of uranium with cancer mortality risk, cancer morbidity risk, and LADD, along with statistical parameters of Bangalore Urban district during pre-monsoon period

Statistical parameters	Uranium [µg/L]	Cancer mortality risk	Cancer morbidity risk	LADD [µg/kg/day]	HQ
Minimum	0.94	3.96.10-8	6.12·10 <sup>-8</sup>	0.06	0.11
Average	23.52	9.91.10-7	1.53.10-6	1.72	2.87
Maximum	98.79	4.16.10-6	6.44·10 <sup>-6</sup>	7.23	12.06
Median	16.47	6.94·10 <sup>-7</sup>	1.07.10-6	1.20	2.01
SD	20.66	8.71.10-7	1.34.10-6	1.51	2.52
Q <sub>25</sub>	8.20	3.45.10-7	5.34.10-7	0.60	1.00
Q <sub>75</sub>	34.654	1.46.10-6	2.25.10-6	2.54	4.23
P <sub>10</sub>	4.10	1.72.10-7	2.67.10-7	0.30	0.50
P <sub>90</sub>	57.51	2.42·10 <sup>-6</sup>	3.74.10-6	4.21	7.02

Explanations: SD – standard deviation;  $Q_{25} - 1^{st}$  quartile;  $Q_{75} - 3^{rd}$  quartile;  $P_{10} - 10^{th}$  percentile;  $P_{90} - 90^{th}$  percentile; LADD – lifetime average daily dose; HQ – hazard quotient.

Table 6. Concentration of uranium with cancer mortality risk, cancer morbidity risk, and LADD, along with statistical parameters of Bangalore Urban district during post-monsoon period

Statistical parameters	Uranium [µg/L]	Cancer mortality risk	Cancer morbidity risk	LADD [µg/kg/day]	HQ
Minimum	1.38	5.81·10 <sup>-8</sup>	8.99.10-8	0.10	0.16
Average	25.60	$1.07 \cdot 10^{-6}$	1.66.10-6	1.87	3.12
Maximum	96.52	4.07·10 <sup>-6</sup>	6.29.10-6	7.07	11.78
Median	19.09	8.05.10-7	1.24.10-6	1.39	2.33
SD	21.30	8.98.10-7	1.38.10-6	1.56	2.60
Q <sub>25</sub>	9.402	3.96.10-7	6.13·10 <sup>-7</sup>	0.68	1.14
Q <sub>75</sub>	33.29	1.40.10-6	2.17.10-6	2.43	4.06
P <sub>10</sub>	4.505	1.89.10-7	2.93.10-7	0.33	0.55
P <sub>90</sub>	50.99	2.15.10-6	3.32.10-6	3.73	6.22

Explanations: SD – standard deviation;  $Q_{25} - 1^{st}$  quartile;  $Q_{75} - 3^{rd}$  quartile;  $P_{10} - 10^{th}$  percentile;  $P_{90} - 90^{th}$  percentile; LADD – lifetime average daily dose; HQ – hazard quotient.

#### **Chemical Risk**

The chemical risk that is caused due to the ingestion of uranium-contaminated water is assessed based on the values of the hazard quotient (HQ) as a non-carcinogenic impact. The chemical hazard that is caused due to uranium ingestion targets the kidneys and lungs [31]. Radiologic toxic effects typically occur at higher exposure levels than chemical toxic effects (which primarily affect the kidneys). The average daily dose of uranium that is consumed throughout one's lifespan was taken into consideration while calculating the chemical toxicity risk, keeping the reference dose (RFD) as 1  $\mu$ g/kg/day as the standard criterion. The hazard quotient and lifetime average daily dose (LADD) can be calculated as shown in Equations (4) and (5) [52]:

$$HQ = \frac{LADD}{RFD}$$
(4)

$$LADD = \frac{EPC \cdot IR \cdot EF \cdot LE}{AT \cdot BW}$$
(5)

where EPC is the exposure point of concentration  $[\mu g/L]$ , IR is the water ingestion rate (taken as 4.05 L/day), EF is the exposure frequency (taken as 350 days), LE is life expectancy (taken as 70.2 years; i.e., 25,703.2 days), AT is the average time [days], and BW is body weight (53 kg). Here, the reference dose can also be referred to as the tolerable daily dose (TDI). Kurttio et al. [30] researched epidemiological studies on consuming water with significant amounts of uranium in it, wherein the WHO (World Health Organization) has prescribed a new guideline value of 1 µg/kg/day as the TDI value. A statistical analysis was performed that analyzed the minimum value, maximum value, average, median (the number in the center of the list of values that are in ascending order), standard deviation (this represents a "typical" variance from the average),  $Q_{25}$  and  $Q_{75}$  (these quantiles are the cut points that divide a probability distribution's range into continuous intervals with equal probabilities; more precisely, they divide a sample's observations into equal groups),  $P_{10}$  ( $P_{10}$  is the value that is greater than 10% of the data points), and  $P_{q_0}$  ( $P_{q_0}$  is the value that is greater than 90% of the data points). During the pre-monsoon season, the LADD values ranged from 0.101 to 7.072  $\mu$ g/kg/day (with an average of 1.876  $\mu$ g/kg/day) as per Table 5, and the HQ (hazard quotient) values varied from 0.16 to 11.78 (with an average value of 3.127). During the post-monsoon season, the LADD values ranged from 0.069 to 7.239  $\mu$ g/kg/day (with an average of 1.72  $\mu$ g/kg/day) as per Table 6, and the HQ values varied from 0.115 to 12.065 (with an average value of 2.872) [28]. When the HQ values were compared accordingly (keeping the RFD value as 1), an apartment at Kaggalipura (Brigade Meadows) witnessed an HQ value of 12.065 during the pre-monsoon season; similarly, Dibbur Village reported an HQ value of 11.78 during the post-monsoon season. These were 11- to 12-times greater than the HQ value of 1 (Fig. 14). These findings demonstrated that the residents of this area were vulnerable to chemical toxicity as a result of consuming uranium. This reference line is implacable to the HQ value of 2.872 that was obtained during the post-monsoon period. By considering the lower cases (P10) as 0.501 and 0.55 during the pre-monsoon and post-monsoon periods, the people were safe from chemical toxicity [23]. The drinking water had an HQ value that was below 1 and was suitable for human consumption; however, the people were 6–7 times more susceptible to chemical poisoning when the upper case (P90) was considered.



Fig. 14. Statistical data of chemical toxicity during pre-monsoon (a) and post-monsoon (b) seasons

# 5. Conclusion

The present study focused on the distribution of uranium in groundwater samples from the Bangalore Urban district as well as its radiological and chemical risk-assessment. According to a data analysis of the current investigation, more than 90% of the water samples were below the acceptable limit of 60  $\mu$ g/L [53]. Very few samples exceeded the AERB limit from either the pre-monsoon or postmonsoon seasons.

The study highlighted the spatial and temporal distribution of uranium as well as its radiological and chemical risk impacts that are caused when its concentration exceeds 60  $\mu$ g/L. This study revealed the chemical toxicity that is caused due to uranium consumption through its values of the hazard quotient (HQ). Hence, consuming water with uranium concentrations of greater than 60  $\mu$ g/L is carcinogenic to human health. Stating its above significance, it is high time to adopt periodic water

analyses and other relevant clinical investigations in the Bangalore Urban district. Hence, the proper monitoring and treatment of high uranium drinking water sources in this area are necessary in order to protect public health, and more research is required to comprehend the hazards to public health that are caused by uranium in the groundwaters in granitic aquifers.

## **Author Contributions**

Poojashri Ravindra Naik: conceptualization, methodology, formal analysis, investigation, data curation, writing – original draft preparation, writing – review and editing.

Vinod Alurdoddi Rajashekara: validation, visualization, review and editing, supervision.

Rajalakshmi Mudbidre: project management, formal analysis, data curation and supervision.

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# References

- [1] Yasovardhan N., Basha A.M., Satyanarayana S.V., Reddy G.V.S., Vishwa K.P., Padma S.P.: Seasonal assessment of natural uranium in drinking water around Tummalapalle Uranium Mining Site, Kadapa District. International Journal of Recent Scientific Research India, vol. 4(9), 2013, pp. 1406–1409.
- [2] Kansal S., Mehra R., Singh N.P.: Uranium concentration in ground water samples belonging to some areas of Western Haryana, India using fission track registration technique. Journal of Public Health and Epidemiology, vol. 3(8), 2011, pp. 352–357.
- [3] Suksi J., Tullborg E.-L., Pidchenko I., Krall L., Sandström B., Kaksonen K., Vitova T., Kvashnina K.O., Göttlicher J.: Uranium remobilisation in anoxic deep rock-groundwater system in response to late Quaternary climate changes – Results from Forsmark, Sweden. Chemical Geology, vol. 584, 2021, 120551. https:// doi.org/10.1016/j.chemgeo.2021.120551.

- [4] Kumar A., Tripathi R.M., Rout S., Mishra M.K., Ravi P.M., Ghosh A.K.: Characterization of groundwater composition in Punjab state with special emphasis on uranium content, speciation and mobility. Radiochimica Acta, vol. 102(3), 2014, pp. 239–254. https://doi.org/10.1515/ract-2014-2109.
- [5] Kumar A., Usha N., Sawant P.D., Tripathi R.M., Raj S.S., Mishra M., Rout S., Supreeta P., Singh J., Kumar S., Kushwaha H. S.: *Risk assessment for natural uranium in subsurface water of Punjab State, India*. Human and Ecological Risk Assessment, vol. 17(2), 2011, pp. 381–393. https://doi.org/10.1080/10807039. 2011.552395.
- [6] Babu M.N.S., Somashekar R.K., Kumar S.A., Shivanna K., Krishnamurthy V., Eappen K.P.: *Concentration of uranium levels in groundwater*. International Journal of Environmental Science and Technology, vol. 5(2), 2008, pp. 263–266. https://doi.org/10.1007/BF03326020.
- [7] Jojo P.J., Rawath A., Ashvani K., Prasad R.: *Trace uranium analysis of water from the south-west coastal region of India*. Journal of Radioanalytical and Nuclear Chemistry, vol. 178(2), 1994, pp. 245–251. https://doi.org/10.1007/ BF02039718.
- [8] Giri S., Singh G., Jha V.N.: Evaluation of radionuclides in groundwater around proposed uranium mining sites in Bagjata and Banduhurang, Jharkhand (India). Radioprotection, vol. 46(1), 2011, pp. 39–57. https://doi.org/10.1051/radiopro/2010056.
- [9] Ramadan R.S., Dawood Y.H., Yehia M.M., Gad A.: Environmental and health impact of current uranium mining activities in southwestern Sinai, Egypt. Environmental Earth Science, vol. 81(7), 2022, 213. https://doi.org/10.1007/s12665-022-10341-9.
- [10] Chaki A., Purohit R.K., Mamallan R.: Low grade uranium deposits of India a bane or boon. Energy Procedia, vol. 7, 2011, pp. 153–157. https://doi.org/10.1016/ j.egypro.2011.06.020.
- [11] Drury J.S., Reynolds S., Owen P.T., Ross R.H., Ensminger J.T.: Uranium in US surface, ground, and domestic waters. Oak Ridge National Lab., Oak Ridge, Tennessee 1981.
- [12] Rowland R.E., Gustafson P.F.: Radiological Physics Division Annual Report: July 1968 through June 1969. Argonne National Laboratory, Argonne 1969. https://rminucleardocs.icaad.ngo/api/files/1610635758681uivrdq6o10m.pdf [access: 20.06.2023].
- [13] Smedley P.L., Kinniburgh D.G.: Uranium in natural waters and the environment : Distribution, speciation and impact. Applied Geochemistry, vol. 148, 2023, 105534. https://doi.org/10.1016/j.apgeochem.2022.105534.
- [14] Carvalho F., Joao M., Margarida M.: Radiological quality of water in areas with old uranium mines in the district of Viseu, Portugal. Journal of Nuclear Energy Science & Power Generation Technology, vol. 4(2), 2015, 1000134. https:// doi.org/10.4172/2325-9809.1000134.

- [15] Juanjuan S., Zhigang Y., Bochao X., Wenhua D., Dong X., Xueyan J.: Concentrations and fluxes of dissolved uranium in the Yellow River estuary: Seasonal variation and anthropogenic (Water-Sediment Regulation Scheme) impact. Journal of Environmental Radioactivity, vol. 128, 2020, pp. 38–46. https:// doi.org/10.1016/j.jenvrad.2013.11.003.
- [16] Sethy N.K., Tripathi R.M., Jha V.N., Sahoo S.K., Shukla A.K., Puranik V.D.: Assessment of natural uranium in the ground water around Jaduguda Uranium Mining Complex, India. Journal of Environmental Protection, vol. 2(7), 2011, pp. 1002–1007. https://doi.org/10.4236/jep.2011.27115.
- [17] Pramod Kumar M., Nagalakshmi K., Jayaraju N., Lakshmi Prasad T., Lakshmanna B.: Deciphering water quality using WQI and GIS in Tummalapalle Uranium Mining area, Cuddapah Basin, India. Water Science, vol. 34(1), 2020, pp. 65–74. https://doi.org/10.1080/11104929.2020.1765450.
- [18] Korake S.R., Jadhao P.D.: Investigation of Taguchi optimization, equilibrium isotherms, and kinetic modeling for cadmium adsorption onto deposited silt. Heliyon, vol. 7(1), 2021, e05755. https://doi.org/10.1016/j.heliyon.2020.e05755.
- [19] Kale A., Bandela N., Kulkarni J.: Spatial distribution and risk assessment of naturally occurring uranium along with correlational study from Buldhana district of Maharashtra, India. Journal of Radioanalytical Nuclear Chemistry, vol. 327(2), 2021, pp. 771–787. https://doi.org/10.1007/s10967-020-07556-0.
- [20] Sharma N., Singh J.: Radiological and chemical risk assessment due to high uranium contents observed in the ground waters of Mansa District (Malwa Region) of Punjab State, India: An area of high cancer incidence. Expo Healh, vol. 8(4), 2016, pp. 513–525. https://doi.org/10.1007/s12403-016-0215-9.
- [21] Sar S.K., Sahu M., Singh S., Diwan V., Jindal M., Arora A.: Assessment of uranium in ground water from Durg District of Chhattisgarh state and its correlation with other quality parameters. Journal of Radioanalytical Nuclear Chemistry, vol. 314(3), 2017, pp. 2339–2348. https://doi.org/10.1007/s10967-017-5587-1.
- [22] Meher P.K., Sharma P., Kumar A., Gautam Y.P., Mishra K.P.: Post monsoon spatial distribution of uranium in water of Alaknanda and Ganges river. International Journal of Radiation Research, vol. 13(1), 2015, pp. 95–99. https:// doi.org/10.7508/ijrr.2015.01.014.
- [23] Kashyap K.K., Ghosh M.K.: Uranium concentration and health risk assessments in groundwater samples taken different location of Korba District, Chhattisgarh, India. Ecology, Environment and Conservation, vol. 28(1), 2022, pp. 171–178. https://doi.org/10.53550/EEC.2022.v28i01s.025.
- [24] Shalumon C.S., Sanu K.S., Thomas J.R., Aravind U.K., Radhakrishnan S., Sahoo S.K., Jha S.K., Aravindakumar C.T.: Analysis of uranium and other water quality parameters in drinking water sources of 5 districts of Kerala in southern India and potability estimation using water quality indexing method. HydroResearch, vol. 4, 2021, pp. 38–46. https://doi.org/10.1016/j.hydres. 2021.04.003.

- [25] Adithya V.S., Chidambaram S., Tirumalesh K., Thivya C., Thilagavathi R., Prasanna M.V.: Assessment of sources for higher uranium concentration in ground waters of the Central Tamilnadu, India. IOP Conference Series: Material Science and Engineering, vol. 121, 2016, 012009. https://doi.org/10.1088/1757-899X/ 121/1/012009.
- [26] Muthamilselvan A.: Anomalous concentration of uranium in groundwater in parts of Salem District, Central Tamil Nadu, India. Advances in Earth and Environmental Science, vol. 1(1), 2020, pp. 1–9. https://doi.org/10.47485/2766-2624.1002.
- [27] Thivya C., Chidambaram S., Tirumalesh K., Prasanna M.V., Thilagavathi R., Adithya V.S., Singaraja C.: Lithological and hydrochemical controls on distribution and speciation of uranium in groundwaters of hard-rock granitic aquifers of Madurai District, Tamil Nadu (India). Environmental Geochemistry and Health, vol. 38(2), 2016, pp. 497–509. https://doi.org/10.1007/s10653-015-9735-7.
- [28] Ben Byju S., Sunil A., Reeba M.J., Christa E.P., Vaidyan V.K., Prasad R., Jojo P.J.: Uranium in drinking water from the south coast districts of Kerala, India. Iranian Journal of Radiation Research, vol. 10(1), 2012, pp. 31–36.
- [29] Langmuir D.: Activity Coefficients of Dissolved Species. [in:] Langmuir D., Aqueous Environmental Geochemistry, Prentice-Hall, Upper Saddle River, New Jersey 1997, pp. 123–192.
- [30] Kurttio P., Komulainen H., Leino A., Salonen L., Auvinen A., Saha H.: Bone as a possible target of chemical toxicity of natural uranium in drinking water. Environmental Health Perspectives, vol. 113(1), 2005, pp. 68–72. https:// doi.org/10.1289%2Fehp.7475.
- [31] Ma M., Wang R., Xu L., Xu M., Liu S.: Emerging health risks and underlying toxicological mechanisms of uranium contamination : Lessons from the past two decades. Environment International, vol. 145, 2020, 106107. https://doi.org/10.1016/ j.envint.2020.106107.
- [32] Gulgundi M.S., Shetty A.: Groundwater quality assessment of urban Bengaluru using multivariate statistical techniques. Applied Water Science, vol. 8(1), 2018, 43. https://doi.org/10.1007/s13201-018-0684-z.
- [33] Government of India, Ministry of Water Resources, Central Ground Water Board: Ground Water Information Booklet: Bangalore Rural District, Karnataka. South Western Region Bangalore. March 2013. https://cgwb.gov.in/old\_ website/District\_Profile/karnataka/2012/BANGALORE\_RURAL-2012.pdf [access: 20.06.2023].
- [34] Lapworth D.J., Brauns B., Chattopadhyay S., Gooddy D.C., Loveless S.E., MacDonald A.M., McKenzie A.A., Muddu S., Nara S.N.V.: *Elevated uranium in drinking water sources in basement aquifers of southern India*. Applied Geochemistry, vol. 133, 2021, 105092. https://doi.org/10.1016/j.apgeochem.2021.105092 [access: 20.06.2023].

- [35] Sekhar M., Tomer S. K., Thiyaku S., Giriraj P., Murthy S., Mehta V.K.: Groundwater level dynamics in Bengaluru City, India. Sustainability, vol. 10(1), 2018, 26. https://doi.org/10.3390/su10010026.
- [36] Prakash M.M., Kaliprasad C.S., Narayana Y.: Studies on natural radioactivity in rocks of Coorg district, Karnataka state, India. Journal of Radiation Research and Applied Sciences, vol. 10(2), 2017, pp. 128–134. https://doi.org/10.1016/ j.jrras.2017.02.003.
- [37] Bhangare R.C., Tiwari M., Ajmal P.Y., Sahu S.K., Pandit G.G.: Laser flourimetric analysis of uranium in water from Vishakhapatnam and estimation of health risk. Radiation Protection and Environment, vol. 36(3), 2013, pp. 128–132.
- [38] Al-Eshaikh M.A., Kadachi A.N., Mansoor Sarfraz M.: Determination of uranium content in phosphate ores using different measurement techniques. Journal of King Saud University – Engineering Science, vol. 28(1), 2016, pp. 41–46. https://doi.org/10.1016/j.jksues.2013.09.007.
- [39] Singh S., Rani A., Mahajan R.K, Walia T.P.S.: Analysis of uranium and its correlation with some physico-chemical properties of drinking water samples from Amritsar, Punjab. Journal of Environmental Monitoring, vol. 5(6), 2003, pp. 917–921. https://doi.org/10.1039/b309493f.
- [40] Amakom C.M., Jibiri N.N.: Chemical and radiological risk assessment of uranium in borehole and well waters in the Odeda Area, Ogun State, Nigeria. International Journal of the Physical Sciences, vol. 5(7), 2010, pp. 1009–1014. https://www.webofscience.com/wos/WOSCC/full-record/000281731100014 [access: 20.06.2023].
- [41] Nagaiah N., Mathews G., Balakrishna K.K.M., Rajanna A.M., Naregundi K.: *Influence of physico-chemical parameters on the distribution of uranium in the ground water of Bangalore, India*. Radiation Protection and Environment, vol. 36(4), 2013, pp. 175–180.
- [42] Tripathi R.M., Sahoo S.K., Mohapatra S., Lenka P., Dubey J.S., Puranik V.D.: Study of uranium isotopic composition in groundwater and deviation from secular equilibrium condition. Journal of Radioanalytical and Nuclear Chemistry, vol. 295(2), 2012, pp. 1195–1200. https://doi.org/10.1007/s10967-012-1992-7.
- [43] Meenakshi, Maheshwari R.C.: Fluoride in drinking water and its removal. Journal of Hazardous Materials, vol. 137(1), 2006, pp. 456–463. https:// doi.org/10.1016/j.jhazmat.2006.02.024.
- [44] Drinking Water Specification (Second Revision). Indian Standard IS 10500:2012, Bureau of Indian Standards, 2012.
- [45] Herschy R.W.: Water Quality for Drinking: WHO Guidelines. [in:] Bengtsson L., Herschy R.W., Fairbridge R.W. (eds.), Encyclopedia of Lakes and Reservoirs, Encyclopedia of Earth Sciences Series, Springer, Dordrecht 2012, pp. 876–883. https://doi.org/10.1007/978-1-4020-4410-6\_184.
- [46] Selvi B.S., Vijayakumar B., Rana B.K., Ravi P.M.: Distribution of natural uranium in groundwater around Kudankulam. Radiation Protection and Environment, vol. 39(1), 2016, pp. 25–29.

- [47] Wang Y., Yu R., Zhu G.: Evaluation of physicochemical characteristics in drinking water sources emphasized on fluoride: A case study of Yancheng, China. International Journal of Environmental Research and Public Health, vol. 16(6), 2019, 1030. https://doi.org/10.3390/ijerph16061030.
- [48] Srinivasamoorthy K., Gopinath M., Chidambaram S., Vasanthavigar M., Sarma V.S.: *Hydrochemical characterization and quality appraisal of groundwater from Pungar sub basin, Tamilnadu, India.* Journal of King Saud University – Science, vol. 26(1), 2014, pp. 37–52. https://doi.org/10.1016/j.jksus.2013.08.001.
- [49] Barbera J., Naibert N., Komperda R., Pentecost T.C.: Clarity on Cronbach' s Alpha Use. Journal of Chemical Education, vol. 98(2), 2021, pp. 257–258. https:// doi.org/10.1021/acs.jchemed.0c00183
- [50] Taber K.S.: The use of Cronbach's Alpha when developing and reporting research instruments in science education. Research in Science Education, vol. 48(6), 2018, pp. 1273–1296. https://doi.org/10.1007/s11165-016-9602-2
- [51] Virk H.S., Jakhu R., Bangotra P.: Natural uranium content in ground waters of Mohaliand Fatehgarh Districts of North Punjab (India) for the assessment of excess cancer risk. Global Journal of Human-Social Science B: Geography, Geo-Sciences, Environmental Science & DisasterManagement, vol. 16(4), 2016, pp. 1–6.
- [52] Kale A., Bandela N., Kulkarni J.: Assessment of chemo-radiological risk of naturally occurred uranium in groundwater from the Beed district, India. Journal of Radioanalytical and Nuclear Chemistry, vol. 323(1), 2020, pp. 151–157. https://doi.org/10.1007/s10967-019-06849-3.
- [53] Sahoo S.K., Jha V.N., Patra A.C., Jha S.K., Kulkarni S.K.: Scientific background and methodology adopted on derivation of regulatory limit for uranium in drinking water – A global perspective. Environmental Advances, vol. 2, 2020, 100020. https://doi.org.10.1016/j.envadv.2020.100020.